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Residential exposure to transportation noise in Denmark and incidence of dementia: national cohort study

Manuella Lech Cantuaria,^{1,2} Frans Boch Waldorff,^{3,4} Lene Wermuth,^{5,6} Ellen Raben Pedersen,¹ Aslak Harbo Poulsen,² Jesse Daniel Thacher,² Ole Raaschou-Nielsen,^{2,7} Matthias Ketzel,^{7,8} Jibrán Khan,^{7,9} Victor H Valencia,⁷ Jesper Hvass Schmidt,^{10,11,12,13} Mette Sørensen^{2,14}

For numbered affiliations see end of the article

Correspondence to:

M L Cantuaria
mlca@mmmi.sdu.dk
(ORCID 0000-0003-3201-9606)

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ABSTRACT

OBJECTIVE

To investigate the association between long term residential exposure to road traffic and railway noise and risk of incident dementia.

DESIGN

Nationwide prospective register based cohort study.

SETTING

Denmark.

PARTICIPANTS

1 938 994 adults aged ≥ 60 years living in Denmark between 1 January 2004 and 31 December 2017.

MAIN OUTCOME MEASURES

Incident cases of all cause dementia and dementia subtypes (Alzheimer's disease, vascular dementia, and Parkinson's disease related dementia), identified from national hospital and prescription registries.

RESULTS

The study population included 1 035 500 participants with incident dementia, and of those, 31 219 received a diagnosis of Alzheimer's disease, 8664 of vascular dementia, and 2192 of Parkinson's disease related dementia. Using Cox regression models, 10 year mean exposure to road traffic and railway noise at the most ($L_{den,max}$) and least ($L_{den,min}$) exposed façades of buildings were associated with a higher risk of all cause dementia. These associations showed a general pattern of higher hazard ratios with higher noise exposure, but with a levelling off or even small declines in risk at higher noise levels. In subtype analyses, both road traffic noise and railway noise were associated with a higher risk of Alzheimer's disease, with hazard ratios of 1.16 (95% confidence interval 1.11 to 1.22) for road $L_{den,max} \geq 65$ dB

compared with < 45 dB, 1.27 (1.22 to 1.34) for road $L_{den,min} \geq 55$ dB compared with < 40 dB, 1.16 (1.10 to 1.23) for railway $L_{den,max} \geq 60$ dB compared with < 40 dB, and 1.24 (1.17 to 1.30) for railway $L_{den,min} \geq 50$ dB compared with < 40 dB. Road traffic, but not railway, noise was associated with an increased risk of vascular dementia. Results indicated associations between road traffic $L_{den,min}$ and Parkinson's disease related dementia.

CONCLUSIONS

This nationwide cohort study found transportation noise to be associated with a higher risk of all cause dementia and dementia subtypes, especially Alzheimer's disease.

Introduction

Dementia is one of the greatest health challenges of the 21st century.¹ With the shifting age distribution of the global population, the number of people with dementia is expected to exceed 130 million by 2050.² In view of the rapid progress of this disease and the increased costs to the health and social care system, the identification of potential modifiable risk factors for dementia have become highly relevant for prevention and management.¹ Besides well established risk factors, such as cardiovascular diseases and unhealthy lifestyle, environmental exposures have received growing attention for a possible role in the development and pathogenesis of dementia.^{1,3}

Transportation noise is considered the second worst environmental risk factor for public health in Europe, only surpassed by air pollution.⁴ About 20% of the European population is exposed to transportation noise above the recommended guideline levels of 55 dB,^{5,6} which is equivalent to noise levels in an open office environment and is 5 dB lower than a regular conversation between people at a distance of 1 m (60 dB).⁷ Epidemiological studies have consistently linked transportation noise to various diseases and health conditions, such as coronary heart disease, obesity, and diabetes.^{6,8} The proposed biological mechanisms for an effect of noise on health are noise induced stress reactions, with activation of the autonomic nervous and the endocrine system and subsequent release of stress hormones, affecting several physiological functions.^{9,10} Exposure to noise during the night can also lead to sleep disturbance and fragmented sleep.¹¹ Experimental studies have found associations between transportation noise at night time and endothelial dysfunction, increased oxidative stress, alterations in the immune system, and increased systemic inflammation,¹²⁻¹⁵ which in turn are regarded as

WHAT IS ALREADY KNOWN ON THIS TOPIC

Environmental exposures have received growing attention for possible roles in the development and pathogenesis of dementia

Previous research on transportation noise and dementia is scarce and diverse

The few existing studies on the topic generally indicated no association between transportation noise and dementia, but some suggested an association with cognitive impairment in adults

WHAT THIS STUDY ADDS

Long term residential exposure to road traffic and railway noise at the most and least exposed façades of buildings was associated with increased risk of all cause dementia and dementia subtypes, especially Alzheimer's disease

The associations between road traffic and all cause dementia followed a positive exposure-response association, with a levelling off or small declines in risk at high noise levels

early events in the onset of dementia and Alzheimer's disease.¹⁶⁻²¹

The scarce existing research on noise and dementia generally does not indicate associations between transportation noise and dementia risk.²²⁻²⁴ However, as concluded by a systematic review in 2020, the current evidence in this research area is of low quality, because of few epidemiological studies with high diversity in study design and outcome definition.²⁵ A few studies have, however, suggested that transportation noise might lead to cognitive impairment in adults.²⁶⁻²⁸ Therefore, further high quality population based studies are required to test the hypothesis of a potential association between exposure to transportation noise and risk of dementia.²⁵

We investigated the association between long term residential exposure to transportation noise and risk of incident dementia in a nationwide cohort including adults aged ≥ 60 years living in Denmark. Register based individual level information was obtained on address history, socioeconomic indicators, and incident all cause dementia and dementia subtypes.

Methods

Study population

This study is based on the population of Denmark, where all residents are followed across national health and administrative registers using a unique personal identification number.²⁹ From the Danish Civil Registration System we identified adults aged ≥ 60 years living in Denmark between 1 January 2004 and 31 December 2017.³⁰ Using this register, we also obtained the histories for residential addresses from 1994 until 2017, to guarantee at least 10 years of address history for participants.³⁰ The date of baseline was either 1 January 2004 or the date at which the participants turned 60 years, whichever came last. We excluded residents with an incomplete address history in the 10 years before baseline, to ensure long term exposure for participants, and those with a dementia diagnosis before baseline. Start of follow-up was chosen as 2004, because 2003 marks a noticeable improvement in the diagnostic rate of dementia compared with previous years.³¹

Identification of outcomes

We identified people with dementia by linking their personal identification numbers to nationwide health registries. All cause dementia was defined as primary or secondary diagnoses of dementia for inpatient and outpatient contacts recorded in the Danish National Patient Register or the Danish Psychiatric Central Register (method 1),³²⁻³³ or at least one prescription of an antidementia drug (donepezil, rivastigmine, galantamine, or memantine) registered in the Danish National Prescription Registry (method 2), or both.³⁴ For subtype analyses, we considered specific diagnoses of Alzheimer's disease, vascular dementia, and Parkinson's disease related dementia. Supplementary table 1 lists the diagnostic codes used for all cause dementia and dementia subtypes. We defined date of

diagnosis as the date of the first recorded dementia diagnosis or the date of the first prescription of an antidementia drug, whichever came first.

Noise exposure assessment

We estimated exposures to road traffic and railway noise at the most and least exposed façades of buildings for all residential addresses in Denmark. Noise levels were calculated as the equivalent continuous A-weighted sound pressure levels (L_{Aeq}) during the day (7 am to 7 pm), evening (7 pm to 10 pm), and night (10 pm to 7 am) and are presented as L_{den} (ie, the average sound level over 24 hours, considering a 5 dB penalty for the evening and a 10 dB penalty for the night, to account for higher susceptibility to noise and stress in these periods).³⁵

The calculations of road traffic and railway noise were based on the Nordic prediction method,³⁶ and Nord2000 models,³⁷ respectively, using the software SoundPLAN (version 8.0, SoundPLAN Nord ApS, Odense, Denmark). Both models are based on the same acoustical principals and have been validated previously—one study that compared measured and predicted values for 174 sites in four Nordic countries (Denmark, Sweden, Norway, and Finland), found an average difference of only 0.3 dB (SD 3.0 dB).³⁸

Major input variables for road traffic noise estimations were geographical address coordinates and height (floor) of each address, and traffic information extracted from a national database on, for example, annual average daily traffic (for 1995, 2000, 2005, 2010, and 2015), vehicle distribution (light, heavy), travel speed, and road type (motorway, rural highway, road > 6 m, road 3-6 m, and other road).³⁹⁻⁴⁰ Major input variables for the railway noise models included geocoded address data, information on annual average daily train lengths, travel speed, and train types obtained from the Danish state railway network. Both road traffic and railway models considered the screening effects from terrains (terrain information retrieved from the GeoDanmark in 2.5 m contour lines), buildings (three dimensional building footprints), noise barriers, and noise berms; ground absorption; and first and second order noise reflections from building façades. We estimated noise exposures at the centre of all façades for each specific address. Buildings with several address points, such as apartments and townhouses, were split into different polygons, resulting in one building polygon for each address point. The lowest and highest noise levels for each address were expressed as exposure at the most ($L_{den,max}$) and least ($L_{den,min}$) exposed façades, respectively.

We modelled road traffic noise for all Danish addresses for 1995, 2000, 2005, 2010, and 2015, and railway noise for all addresses within 1000 m of a rail track for 1997 and 2012. Based on the estimations for these five time points for road traffic noise and two time points for railway noise, we applied linear interpolation to calculate yearly noise estimates for each of the years from 1994 to 2017 for all Danish addresses.

We subsequently linked the address specific yearly noise exposures at all Danish residential addresses to the person specific address histories obtained for the whole study population.

For each cohort member we then calculated time weighted 10 year running means for $L_{den,max}$ and $L_{den,min}$, taking exposure at all addresses in the periods into account. The time weighted means were calculated as arithmetic means on the sound pressure scale followed by logarithmic transformation to the decibel scale. All noise values below 35 dB were then set to 35 dB because we do not expect people to perceive transportation noise levels below this threshold because of background noise from other sources.

Covariates

We collected a variety of covariates from registers available at Statistics Denmark: yearly individual level variables (from 2004 to 2017) on civil status (married or cohabiting, widowed, divorced, and single), country of origin (Denmark, other western country, non-western country), individual income (fifths, based on the yearly distribution among Danish adults aged 25-70 years), occupational status (blue collar, low level white collar, high level white collar, unemployed, retired), and highest attained education (mandatory, secondary or vocational, medium or long education). Furthermore, for each of the 2160 parishes available in Denmark, we gained information on population density (inhabitants/km²) and neighbourhood level socioeconomic status (ie, the proportion of residents with low income (lowest fourth), unemployed, in manual labour, with only basic education, with a criminal record, and living in a single parent household). We also extracted information on the building type (detached house, semidetached house, multistorey building, and other) for each address from the Building and Housing Register.

We calculated the proportion of high quality green space (ie, recreational areas, forests, and wet and dry nature areas) within 150 m radius and 1000 m radius of each address, using land use and land cover maps of Denmark.⁴¹ Furthermore, for all Danish addresses, we estimated concentrations of fine particulate matter (PM_{2.5}) and nitrogen dioxide for 2000, 2010, and 2015 using the Danish DEHM-UBM-AirGIS modelling system. Calculations in DEHM-UBM-AirGIS combine three models operating at different spatial scales (DEHM: regional background, UBM: urban background, and AirGIS: traffic in nearest streets), taking into account factors such as road and traffic information, street configurations, emissions from all Danish sources in high resolution, and background concentrations from outside Denmark.⁴² We subsequently calculated yearly and spatially averaged urban background PM_{2.5} and nitrogen dioxide levels in Denmark for 1994-2017 using the DEHM-UBM model. These yearly background means were used to calculate yearly scaling factors, which we used to extrapolate the address specific estimates from 2000, 2010, and 2015 into yearly address specific means for 1994-2017.

Statistical analysis

Cox proportional hazard models were used with age as the underlying time to calculate hazard ratios for the association between transportation noise at the most exposed façades ($L_{den,max}$) and least exposed façades ($L_{den,min}$) and all cause dementia and dementia subtypes. We right censored at the age of a dementia diagnosis, death, missing address, emigration, or end of follow-up (31 December 2017), whichever came first. The time weighted running means for road traffic and railway noise were included as time varying variables into the Cox model; thus for each participant with a dementia diagnosis comparing with the 10 year mean exposure for all participants without a dementia diagnosis at the same age as the participant with dementia.

To test the assumption of proportional hazards for all covariates, we evaluated the independence between scaled Schoenfeld residuals with time. We included sex, civil status, occupational status, income, and calendar year as strata, as these deviated from the assumption of proportional hazards.

We analysed the associations between 5 dB categories of 10 year mean road traffic and railway $L_{den,max}$ and $L_{den,min}$ and all cause dementia using three models: model 1, adjusted for age (by design), sex, and calendar year; model 2, with further adjustment for region of origin, civil status, personal income, occupational status, proportion of high quality green space, and area level socioeconomic variables, as well as mutual adjustment for road traffic and railway noise; and model 3, with additional adjustment for time weighted 10 years running means of PM_{2.5}. We used <45 dB as the reference category for all analyses with road traffic $L_{den,max}$, and <40 dB for all analyses with road traffic $L_{den,min}$ and railway $L_{den,max}$ and $L_{den,min}$. Associations were also analysed between categories of 10 year mean exposure to road traffic and railway noise and dementia subtypes (Alzheimer's disease, vascular dementia, and Parkinson's disease related dementia) adjusting for the covariates in model 2.

We chose the covariates for the statistical model based on availability in national registers and a literature review of plausible mechanisms behind a potential association between transportation noise and dementia, as shown by the directed acyclic graph (supplementary figure 1). A priori we selected model 2 as our main adjustment model. All covariates, apart from sex and region of origin, were included in the statistical models as time varying variables, therefore allowing for change throughout the follow-up period.

Natural cubic splines with four degrees of freedom (based on adjustment of model 2) were applied to investigate the shape of the exposure-response association between road traffic and railway noise ($L_{den,max}$ and $L_{den,min}$) and risk of all cause dementia, and between road traffic noise ($L_{den,max}$ and $L_{den,min}$) and dementia subtypes (Alzheimer's disease, vascular dementia, and Parkinson's disease related dementia).

The joint effect of exposure to road traffic and railway noise in relation to all cause dementia was

investigated by combining three categories of road traffic noise (L_{den} max of <55 dB, 55-65 dB, and \geq 65 dB, and L_{den} min of <45 dB, 45-55 dB, and \geq 55 dB) and three categories of railway noise (L_{den} max of <40 dB, 40-50 dB, and \geq 50 dB, and L_{den} min of <35 dB, 35-45 dB, and \geq 45 dB) using low road traffic and railway noise as reference. Similarly, we investigated the joint effect of exposure to road traffic noise at the most and least exposed building façades in relation to all cause dementia by combining three categories of L_{den} max (<55 dB, 55-60 dB, and \geq 60 dB) and three categories of L_{den} min (<40 dB, 40-50 dB, and \geq 50 dB) using low L_{den} max and L_{den} min as reference.

For all cause dementia, we conducted sensitivity analyses based on model 2 by including both L_{den} max and L_{den} min; adjusting for both time weighted exposures to $PM_{2.5}$ and nitrogen dioxide; further adjusting for time varying highest attained education for the subpopulation for which this information was available; and including only cases with a primary dementia diagnosis. We also investigated associations with categories of 10 year time weighted night time noise (L_{night}).

We evaluated possible effect modification of the association between 10 year mean road traffic noise at the least exposed façade and all cause dementia by generating natural cubic splines with four degrees of freedom (based on adjustment of model 2) for each of the following groups: men and women; low, medium, and high population density; detached house, semidetached house, and multistorey building; low, medium, and high income; before and after 2010; and with and without comorbidity from myocardial infarction, stroke, or diabetes.

All statistical analyses were performed in SAS 9.4 (SAS Institute, Cary, NC), apart from the analysis of proportional hazards and natural splines, for which we used R, version 3.6.3 (survival package).

Patient and public involvement

As this is a register based study, no patients or members of the public were involved in the design, conduct, reporting, or dissemination plans of our research.

Results

The study base included 2.1 million Danish residents aged \geq 60 years. People were excluded if they had a diagnosis of dementia before baseline ($n=87\,109$), had an incomplete history on exposure before enrolment ($n=69\,692$), or were missing covariates ($n=2411$), resulting in a study population of 1 938 994 participants and 103 500 participants with incident dementia. The mean follow-up was 8.5 years (ie, a total of 16 425 922 person years), with an event rate of 6.3 cases/1000 person years. Of the participants with dementia, 18 628 were exclusively identified by method 2 (supplementary figure 2), and 1899 (<2% of participants with dementia) were identified based on less than three prescriptions for antedementia drugs. In total, 31 219 participants received a diagnosis of Alzheimer's disease, 8664 a diagnosis of vascular

dementia, and 2192 a diagnosis of Parkinson's disease related dementia. The incidence rate of dementia in the study population was stable during the study period, with, for example, 7282 participants identified with dementia in 2004 and 7058 participants identified with dementia in 2017.

Table 1 and supplementary table 2 show the baseline characteristics of the study population for road traffic L_{den} max below and above 55 dB and for participants with or without dementia. The median 10 year exposure was 55.3 dB (L_{den} max) and 44.3 dB (L_{den} min) for road traffic noise, and 51.6 dB (L_{den} max) and 44.7 dB (L_{den} min) for railway noise (among those who were exposed). Supplementary figures 3 and 4 show the distribution of exposure to road traffic and railway noise at baseline, and supplementary figure 5 shows the distribution of the difference between noise measured at the most and least exposed façades.

For road L_{den} max, Spearman correlation coefficients of 0.49 were found for road L_{den} min, 0.12 for railway L_{den} max, 0.10 for railway L_{den} min, 0.25 for $PM_{2.5}$, and 0.45 for nitrogen dioxide. Only weak correlations were observed between these variables and proportion of high quality green space (supplementary table 3). Results also showed high correlations between L_{den} and L_{night} , especially for road traffic noise (about 1.00 for most and least exposed façades), but also for railway noise (0.94 for L_{den} max and 0.83 for L_{den} min).

Road traffic and railway noise at the most and least exposed façade were associated with increased risk of all cause dementia across all exposure categories (table 2) with, for example, hazard ratios of 1.16 (95% confidence interval 1.13 to 1.19) for road L_{den} max \geq 65 dB (compared with <45 dB), 1.18 (1.15 to 1.21) for road L_{den} min \geq 55 dB (compared with <40 dB), 1.16 (1.12 to 1.20) for railway L_{den} max \geq 60 dB (compared with <40 dB), and 1.16 (1.12 to 1.19) for railway L_{den} min \geq 50 dB (compared with <40 dB). The associations for road traffic noise followed exposure-response associations until about 55 dB for L_{den} max and 50 dB for L_{den} min, after which no further increase or even a slight decrease in risk was observed (fig 1). Hazard ratios were slightly higher for L_{den} min compared with L_{den} max. Similarly, a positive exposure-response association was found for railway noise L_{den} max, with a slight decline in risk at higher exposures (fig 1). For railway L_{den} min, however, an upward trend was observed until about 45 dB, with a levelling off in risk at medium exposure levels (45 dB to 55 dB), followed by an increase in risk from 55 dB. Associations between exposure to noise at night time (L_{night}) and all cause dementia were generally similar to the associations observed for L_{den} (supplementary table 4).

Based on the hazard ratios and the percentage of exposed people in each noise category, the population attributable risk for road traffic and railway noise was calculated. Of the 8475 patients with incident dementia registered in Denmark in 2017,⁴³ the diagnosis in an estimated 963 patients (95% confidence interval 762 to 1174) was attributed to road traffic noise and in 253 patients (195 to 310) to railway noise.

Table 1 | Characteristics of study population at baseline (n=1 938 994) according to exposure to road traffic noise at most exposed façade of building. Values are numbers (percentages) unless stated otherwise

Characteristics	Exposure level	
	<55 dB (n=946 609)	≥55 dB (n=992 385)
Individual level		
Men	449 413 (47.5)	458 578 (46.2)
Mean (SD) age (years)	65.8 (8.2)	66.5 (8.8)
Country of origin:		
Denmark	937 329 (99.0)	976 742 (98.4)
Other western country	5310 (0.6)	7328 (0.7)
Non-western country	3970 (0.4)	8315 (0.9)
Civil status:		
Married or cohabiting	695 872 (73.5)	644 905 (65.0)
Widowed	124 856 (13.2)	152 267 (15.3)
Divorced	74 634 (7.9)	118 670 (12.0)
Single	51 247 (5.4)	76 543 (7.7)
Individual income fifth*:		
1st (lowest)	260 210 (27.5)	305 196 (30.7)
2nd	219 369 (23.1)	247 982 (25.0)
3rd	149 286 (15.8)	151 689 (15.3)
4th	147 421 (15.6)	134 824 (13.6)
5th (highest)	170 323 (18.0)	152 694 (15.4)
Occupational status:		
Blue collar	85 731 (9.1)	75 598 (7.6)
White collar	346 480 (36.6)	336 163 (33.9)
Unemployed	36 472 (3.8)	43 324 (4.4)
Retired	477 926 (50.5)	537 300 (54.1)
Highest attained education:		
Mandatory education	343 531 (36.3)	375 759 (37.9)
Secondary or vocational education	373 875 (39.5)	373 290 (37.6)
Medium or long education	166 392 (17.6)	158 986 (16.0)
Unknown	62 811 (6.6)	84 350 (8.5)
Address level†		
Proportion of high quality green space:		
≥15% in 150 m radius	207 299 (21.9)	201 448 (20.3)
≥20% in 1000 m radius	259 355 (27.4)	239 881 (24.2)
Urbanity (inhabitants/km ²):		
Low (≤100)	248 130 (26.2)	252 827 (25.5)
Medium (101-4999)	673 684 (71.2)	640 993 (64.6)
High (≥5000)	24 795 (2.6)	98 565 (9.9)
Building type:		
Detached house	633 563 (66.9)	500 529 (50.4)
Semidetached house	162 122 (17.1)	141 659 (14.3)
Multistorey building	149 096 (15.8)	347 269 (35.0)
Other	1828 (0.2)	2928 (0.3)
Mean (SD) 10 year mean air pollution exposure (µg/m ³):		
PM _{2.5}	15.2 (1.7)	15.9 (2.1)
Nitrogen dioxide	16.5 (5.2)	21.3 (8.7)
Mean (SD) area level factors (% of population):		
Low income (1st fourth)	4.4 (2.0)	5.0 (2.4)
Unemployed	1.7 (0.7)	1.8 (0.7)
Manual labour	13.1 (3.1)	12.2 (3.3)
Only basic education	10.6 (3.4)	10.2 (3.5)
Criminal record	0.4 (0.3)	0.5 (0.3)
Single parent families	5.6 (1.7)	5.7 (1.8)

PM_{2.5}=fine particulate matter.

*Standardised by calendar year and sex.

†Based on residential address of participants at baseline.

Higher risk estimates for Alzheimer's disease were found for both road traffic and railway noise, with the highest risk estimates for road traffic L_{den} min (hazard ratio 1.30, 95% confidence interval 1.25 to 1.36) for the exposure category 50-55 dB and 1.27 (1.22 to 1.34) for ≥55 dB, compared with <40 dB; table 2). No clear associations were found between exposure to railway noise and vascular dementia. Furthermore,

no associations were observed between road traffic L_{den} max and risk of Parkinson's disease related dementia, whereas some indication of an association was found for road traffic L_{den} min. However, only a few participants had Parkinson's disease related dementia and therefore confidences were wide for this outcome. Exposure-response associations for road L_{den} min and all dementia subtypes followed an upward trend until about 55 dB for Alzheimer's disease, 50 dB for vascular dementia, and 53 dB for Parkinson's disease related dementia. After these thresholds, we observed a decrease in hazard ratios, however with wide confidence intervals (fig 2). Splines for L_{den} max showed increases in risk for all subtypes until approximately 55 dB for Alzheimer's disease, and 60 dB for vascular dementia and Parkinson's disease related dementia, after which a levelling off in risk was observed.

When investigating associations between categories combining road traffic L_{den} max and L_{den} min in relation to all cause dementia, exposure-response associations were observed for L_{den} min in all exposure categories of L_{den} max, whereas the opposite was not observed (supplementary table 5). In an analysis combining categories of road traffic and railway noise, co-exposure to medium-high levels of both road traffic and railway noise was associated with high hazard ratios—for example, for road traffic L_{den} max ≥65 dB and railway L_{den} max ≥50 dB the hazard ratio was 1.22 (95% confidence interval 1.18 to 1.27; supplementary table 6).

Adjustment for air pollution (PM_{2.5} alone and together with nitrogen dioxide) and education resulted in some reductions in continuous hazard ratios for all cause dementia, although associations with both road traffic and railway noise remained (supplementary tables 7 and 8). Mutual adjustment for L_{den} max and L_{den} min was not associated with substantial changes in risk estimates (supplementary table 10). When restricting analyses to participants with a primary diagnosis of dementia (n=42 413), the hazard ratios for road traffic and railway L_{den} min categories were observed to increase—for example, for the highest exposure category (≥55 dB for road traffic and ≥50 dB for railway) compared with the reference group increased from 1.18 (1.15 to 1.21) to 1.30 (1.25 to 1.35) for road traffic noise and from 1.16 (1.12 to 1.19) to 1.21 (1.16 to 1.27) for railway noise (supplementary table 8). No substantial increase in hazard ratios was, however, observed for railway and road traffic L_{den} max.

Figure 3 presents splines for the associations between 10 year mean road traffic noise (L_{den} min) and all cause dementia in relation to potential effect modifiers. Stronger associations were found between road traffic noise and dementia among people living in low compared with medium and high populated areas, and among people living in detached houses versus semidetached houses and multistorey buildings. Even though the curves overlapped, hazard ratios were found to be slightly lower for people with higher income compared with those with lower or medium income. No substantial differences were observed

Table 2 | Associations between categories of 10 year mean residential exposure to road traffic and railway noise at the most ($L_{den\ max}$) and least ($L_{den\ min}$) exposed façades of buildings and risk of all cause dementia and different dementia subtypes

Noise exposure (10 year) by category	All cause dementia		Alzheimer's disease		Vascular dementia		Parkinson's disease dementia	
	No of cases	Hazard ratio (95% CI)*	No of cases	Hazard ratio (95% CI)*	No of cases	Hazard ratio (95% CI)*	No of cases	Hazard ratio (95% CI)*
Road								
$L_{den\ max}$ (dB):								
<45 (reference)	9718	1	2799	1	827	1	242	1
45-50	13 500	1.09 (1.06 to 1.12)	4081	1.10 (1.04 to 1.15)	1090	1.03 (0.94 to 1.13)	287	0.94 (0.79 to 1.12)
50-55	21 738	1.16 (1.13 to 1.19)	6798	1.19 (1.14 to 1.25)	1754	1.10 (1.01 to 1.20)	463	1.03 (0.88 to 1.21)
55-60	24 533	1.17 (1.14 to 1.20)	7427	1.17 (1.12 to 1.22)	2084	1.17 (1.08 to 1.27)	519	1.07 (0.92 to 1.25)
60-65	21 196	1.16 (1.13 to 1.19)	6242	1.14 (1.09 to 1.19)	1802	1.18 (1.08 to 1.28)	443	1.10 (0.94 to 1.29)
≥65	12 815	1.16 (1.13 to 1.19)	3872	1.16 (1.11 to 1.22)	1107	1.19 (1.09 to 1.31)	238	1.04 (0.87 to 1.25)
$L_{den\ min}$ (dB)								
<40 (reference)	18 136	1	4902	1	1475	1	394	1
40-45	30 306	1.12 (1.10 to 1.14)	8608	1.10 (1.06 to 1.14)	2621	1.19 (1.12 to 1.27)	619	1.10 (0.96 to 1.25)
45-50	27 261	1.18 (1.16 to 1.20)	8403	1.22 (1.17 to 1.26)	2239	1.20 (1.12 to 1.28)	580	1.18 (1.03 to 1.35)
50-55	18 005	1.21 (1.19 to 1.24)	5985	1.30 (1.25 to 1.36)	1567	1.31 (1.21 to 1.41)	402	1.26 (1.08 to 1.46)
≥55	9792	1.18 (1.15 to 1.21)	3321	1.27 (1.22 to 1.34)	762	1.14 (1.04 to 1.25)	197	1.16 (0.96 to 1.39)
Railway								
$L_{den\ max}$ (dB):								
<40 (reference)	75 690	1	22819	1	6409	1	1603	1
40-45	4441	1.09 (1.06 to 1.12)	1287	1.02 (0.96 to 1.08)	387	1.14 (1.03 to 1.26)	117	1.43 (1.18 to 1.72)
45-50	5844	1.11 (1.08 to 1.14)	1737	1.06 (1.01 to 1.12)	488	1.10 (1.00 to 1.20)	122	1.17 (0.97 to 1.41)
50-55	8040	1.19 (1.16 to 1.22)	2491	1.20 (1.15 to 1.26)	656	1.13 (1.04 to 1.23)	149	1.16 (0.98 to 1.38)
55-60	5577	1.19 (1.16 to 1.23)	1695	1.20 (1.15 to 1.27)	403	0.99 (0.90 to 1.10)	134	1.49 (1.25 to 1.78)
≥60	3908	1.16 (1.12 to 1.20)	1190	1.16 (1.10 to 1.23)	321	1.09 (0.97 to 1.22)	67	1.07 (0.84 to 1.37)
$L_{den\ min}$ (dB):								
<40 (reference)	86 286	1	25808	1	7309	1	1845	1
40-45	6049	1.12 (1.09 to 1.15)	1866	1.12 (1.06 to 1.17)	505	1.10 (1.00 to 1.20)	149	1.39 (1.17 to 1.64)
45-50	6302	1.17 (1.14 to 1.20)	1973	1.20 (1.15 to 1.26)	467	1.01 (0.92 to 1.11)	102	0.98 (0.80 to 1.20)
≥50	4863	1.16 (1.12 to 1.19)	1572	1.24 (1.17 to 1.30)	383	1.03 (0.93 to 1.14)	96	1.19 (0.96 to 1.47)

*Model 2 adjusted for age (by design), sex, calendar year, civil status, income, region of origin, occupational status, proportion of high quality green space, and a number of area level socioeconomic variables: percentage population with low income, only basic education, unemployment, manual labour occupation, single parent, and a criminal record, as well as mutual adjustment for road traffic and railway noise. All covariates, apart from sex and region of origin, were included in the model as time varying variables.

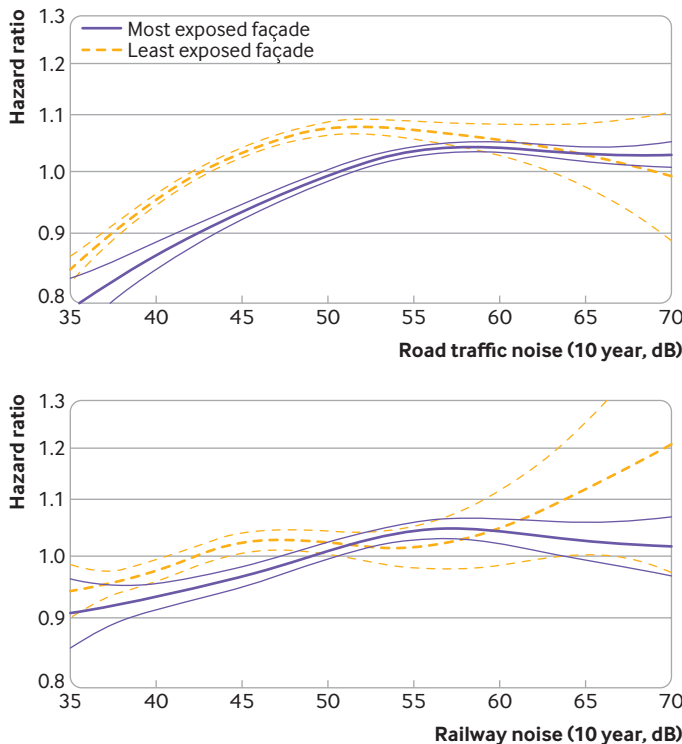


Fig 1 | Associations between 10 year mean exposure to road traffic and railway noise at the most ($L_{den\ max}$) and least ($L_{den\ min}$) exposed façades of buildings and risk of all cause dementia, using the fully adjusted model (model 2). Figure shows hazard ratios and corresponding 95% confidence intervals

in the exposure-response curves in relation to sex, comorbidity, or calendar year. In a subanalysis, positive associations were observed between road traffic noise and all cause dementia for all five regions of Denmark (supplementary figure 6).

Discussion

In our large nationwide cohort, we found long term residential exposure to road traffic and railway noise to be associated with increased risk of all cause dementia. These associations followed positive exposure-response associations, with a levelling off or slight decreases in risk at high noise exposures for most of the exposure metrics. For Alzheimer's disease we observed associations with both road traffic and railway noise, whereas for vascular dementia we only observed clear associations for road traffic noise. We found indications of slightly higher risk estimates for 10 year mean exposure to road traffic noise at the least exposed façades of buildings ($L_{den\ min}$) compared with most exposed ($L_{den\ max}$) for both Alzheimer's disease and vascular dementia. The association between road traffic noise and dementia differed in relation to population density, building type, and income.

Strengths and limitations of this study

One of the major strengths of our study is the nationwide prospective design, which included a large population, long follow-up, and high quality assessment of noise

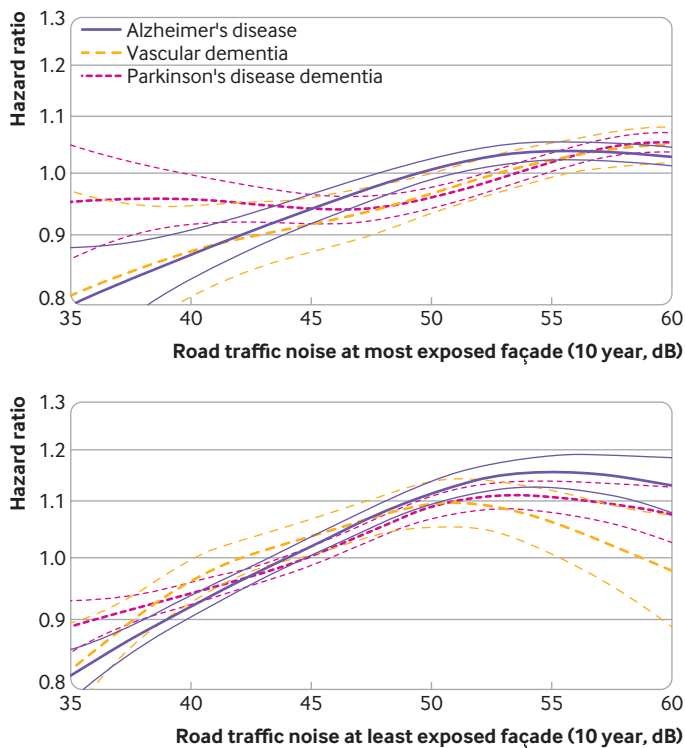


Fig 2 | Associations between 10 year mean exposure to road traffic noise at the most and least exposed façades of buildings and risk of dementia subtypes, using the fully adjusted model (model 2). Figure shows hazard ratios and corresponding 95% confidence intervals

exposure from two different transportation sources. We had access to complete residential address histories and detailed information on several potential confounders, including individual and area level socioeconomic variables, proportion of high quality green spaces, and exposure to air pollutants estimated from state-of-the-art models.⁴² Another strength was the robustness of the noise exposure assessment, which also enabled us to assess noise both at the most and at the least exposed façades of buildings. Furthermore, our study combined data from three unique national registers, allowing for the identification of many participants with all cause dementia with demonstrated high validity.⁴⁴

Limitations include lack of information on lifestyle factors, which are recognised risks for dementia.¹ However, our models included several socioeconomic variables, which are highly associated with lifestyle.⁴⁵ Besides, adjustment for lifestyle could result in overadjustment, as important lifestyle variables (eg, smoking, diet, alcohol consumption, and physical activity) were shown to be potentially associated with transportation noise,^{46 47} thus acting as mediators on the pathway rather than just confounders. Furthermore, we lacked information on factors that might affect personal exposure to noise, such as sound insulation in homes, which increases the risk for exposure misclassification. Also, although the Nordic prediction method has shown good predictive performance,³⁸ all exposure models are inevitably associated with misclassification of exposure. This can be caused by, for example, inaccuracy in input data on traffic

counts and distribution, and lack of information on other noise sources. Such misclassification is probably non-differential, which would tend to drive the risk estimates towards 1. Also, we lacked information on noise sources that might mask the association between road traffic and railway noise and dementia, such as airports and construction sites.

Although we used three high quality nationwide registries to identify participants with dementia, we cannot rule out that outcome misclassification was a problem, mainly for definition of subtypes. One study found a low validity for register based diagnosis of dementia subtypes in Denmark, apart from Alzheimer's disease, and therefore these results should be interpreted with caution.⁴⁴ Moreover, identification of participants with dementia based entirely on prescription of antidementia drugs (method 2) has not been previously validated. Diagnosing dementia is a complicated and non-standardised procedure, with potential differences in patients' socioeconomic status, and regional differences in diagnosis rates and services.^{31 48 49} In a subanalysis, however, we found consistent associations between road traffic noise and all cause dementia for the five regions of Denmark, apart from slightly lower hazard ratios for the capital region, suggesting that regional differences in the diagnosis of dementia did not drive the observed associations in the present study.

Comparison with other studies and discussion of main findings

We found consistent associations between transportation noise from two independent noise sources (road traffic and railway) and risk of dementia. These associations became stronger when only participants with a primary diagnosis of dementia were considered. Previous research in this area is scarce and diverse, as pointed out by a recent systematic review on transportation noise and mental health.²⁵ To the extent of our knowledge, only three studies investigated transportation noise and the incidence of dementia. The first study was based on a cohort of about 130 000 adults living in London and identified participants with all cause dementia, Alzheimer's disease, and vascular dementia over a nine year period through medical records.²³ The second study used a sample of about 678 000 Canadian residents and identified cases of non-Alzheimer's disease dementia and Alzheimer's disease in three administrative health databases.²⁴ That study, however, found a low number of participants with Alzheimer's disease (6.5%), suggesting that outcome misclassification might have been a problem. The third study was based on a population of 1721 adults in Sweden, with assessment of dementia (Alzheimer's disease and vascular dementia) using a three phase procedure of testing, interview, and examination by a specialist.²² In all the studies, researchers found no indications of an association between transportation noise and risk of dementia. In contrast with our study, where we had exact information on the address location

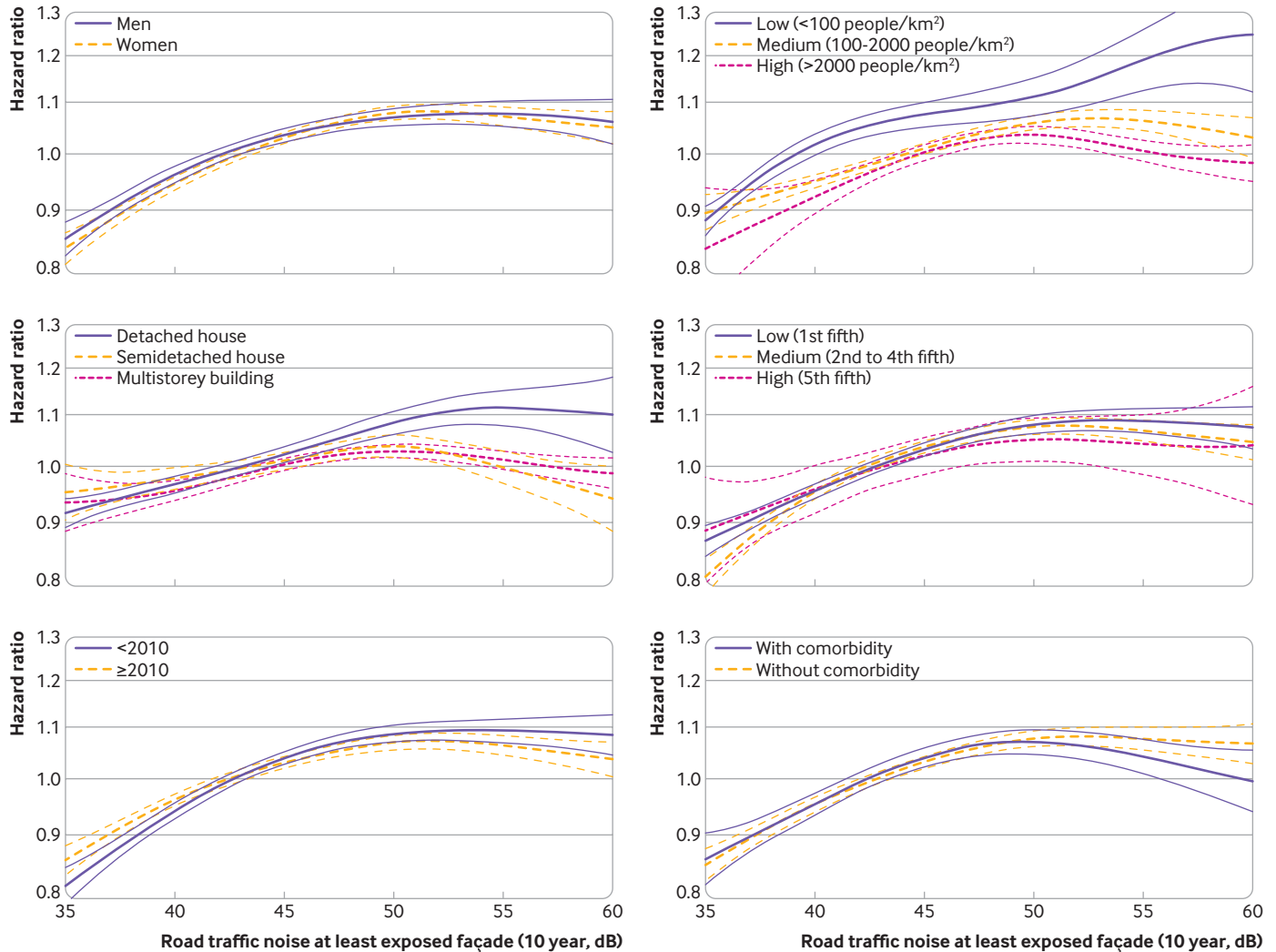


Fig 3 | Associations between 10 year mean exposure to road traffic noise at the least exposed façade of buildings and risk of incident all cause dementia according to sex, population density, type of building, income, calendar year, and comorbidity. Figure shows hazard ratios and corresponding 95% confidence intervals

(including floor) of all participants, the English and Canadian studies estimated road traffic noise levels at postal code level,^{23 24} which increases the chance of exposure misclassification.⁵⁰ The Swedish study used dichotomised exposure levels, reducing variability in the exposure assessment.²² Inaccuracy on the noise estimates in these studies might therefore have driven the risk estimates towards the null. A cross sectional German study of 2050 adults, however, did find positive associations between road traffic noise and overall mild cognitive impairment and amnesic mild cognitive impairment, assessed by five neuropsychological subtests.²⁷ In support, a small study (n=1612) in the US found positive associations between road traffic noise and cognitive impairment or dementia, especially at noise levels ≥ 75 dB.²⁸ Both of these studies had the advantage of relying on residence specific assessment of noise exposure. However, similar to the study in Sweden²² and London,²³ the researchers only estimated noise at the baseline address, therefore not taking residential mobility and

long term exposure before diagnosis into account, as we did in our study.

The *Lancet* Commission report on dementia recently added air pollution as a modifiable risk factor for dementia,¹ owing to the amount of extant research showing air pollution to contribute to an increased risk of dementia.^{3 51 52} Air pollution and transportation noise share a common source (ie, transportation), and therefore including high quality estimations of air pollution in our models is important to disentangle the effect of the two exposures.⁵³ Our results were robust to adjustment for air pollution, thus indicating an independent effect of noise on the biological mechanisms leading to dementia.

Most studies on transportation noise and health have assessed noise only at the most exposed façade of a building. In our study we also assessed exposure at the least exposed façade, which we hypothesise to better describe noise exposure during sleep, given that bedrooms are often located at quiet sides of the buildings.⁵⁴ Several studies have suggested that night

time noise is crucial to the adverse health effects of noise, especially from interrupted sleep, which is a critical period for mental and cognitive restoration.^{14 55} Furthermore, an increasing body of evidence suggests associations between sleep disturbances, such as sleep quality and duration, with increased risk of all cause dementia and Alzheimer's disease.¹ In our study, we found essentially the same risk estimates for whole day noise and only night time noise because of high correlations between both metrics. However, our results indicated slightly higher risk estimates for road traffic $L_{den, min}$ compared with $L_{den, max}$, especially for Alzheimer's disease and vascular dementia, but also for all cause dementia. This was supported by the analysis of combined effects of $L_{den, max}$ and $L_{den, min}$, which showed exposure-response associations for $L_{den, min}$ in all exposure categories of $L_{den, max}$, whereas the opposite was not observed. This suggests that effects of noise on sleep might play an important role in increasing risk of dementia, which corroborates with previous studies on sleep quality and dementia.^{1 16-18} Also, one study found more pronounced associations between night time noise and mild cognitive impairment compared with exposure to noise during the whole day.²⁷

When we investigated the association between noise and dementia subtypes, we found the highest and most consistent risk estimates for Alzheimer's disease. Although much is to be explored about the underlying mechanisms in Alzheimer's disease, several animal studies have shown chronic exposure to noise to exacerbate age related changes and activate the formation of key role genes that potentially act as mediators for pathological changes related to Alzheimer's disease in the hippocampus of mice.⁵⁶⁻⁵⁹ Because Parkinson's disease related dementia shares similar pathology to Alzheimer's disease (eg, accumulation of amyloid β plaques and formation of tau hyperphosphorylation in the brain), and both diseases can often coexist,⁶⁰ we expected compatible results for both Alzheimer's disease and Parkinson's disease related dementia. Nevertheless, results were not so consistent, especially for railway noise.

Previous studies found transportation noise to be associated with increased risk of stroke,⁶⁸ which corroborates our results of an association between road traffic noise and vascular dementia, as vascular dementia is caused by poor blood flow to the brain.⁶¹ In agreement with our findings, a recent nationwide Danish study found road traffic noise to increase the risk of incident stroke, whereas no clear associations were observed for railway noise.⁶² Studies on exposure to transportation noise and stroke are not, however, consistent,^{63 64} thus indicating a need for further studies.

For most of the associations, we observed a levelling off or decline in hazard ratios at high levels of noise, especially for road traffic noise, potentially related to investments in better sound insulation at higher noise levels.^{54 65} In support, we observed lower risk estimates among residents living in medium and highly populated areas and living in the capital region,

which we would expect to be groups prioritising such measures. In line with our findings, previous studies on transportation noise and mortality also estimated higher risks in suburban areas compared with urban areas.^{50 66} Furthermore, exposure misclassification is expected to be higher in densely populated areas, as inner city areas with narrow streets and high buildings tend to have higher estimation error as a result of many reflections, and our model includes only first and second order reflections. Besides, traffic congestion is more common in highly populated areas, which results in higher noise levels not accounted for by the noise model. Lastly, the presence of many competing noise sources in highly populated areas from, for example, community life, might lead to masking of the association between transportation noise and dementia.⁵⁰ In line with this, we found lower hazard ratios for residents living in multistorey buildings and semidetached houses compared with detached houses. Other potential explanations for this levelling off might be a potential survivor effect, as residents who are less tolerant to noise might have moved away from highly exposed areas or died from competing diseases related to noise, such as cardiovascular diseases. At present, however, it is unclear whether people with a low tolerance to noise are more susceptible to health effects of noise, as findings on effect modification by noise annoyance are inconsistent.⁶⁷⁻⁶⁹

Conclusions

In this large nationwide cohort study, we found transportation noise from road traffic and railway to be associated with increased risk of all cause dementia and dementia subtypes, especially Alzheimer's disease. If these findings are confirmed in future studies, they might have a large effect on the estimation of the burden of disease and healthcare costs attributed to transportation noise. Expanding our knowledge on the harmful effects of noise on health is essential for setting priorities and implementing effective policies and public health strategies focused on the prevention and control of diseases, including dementia.

AUTHOR AFFILIATIONS

¹The Mærsk McKinney Møller Institute, University of Southern Denmark, Odense, Denmark

²Diet, Genes and Environment, Danish Cancer Society Research Center, Copenhagen, Denmark

³Department of Public Health, The Research Unit for General Practice and Section of General Practice, University of Copenhagen, Copenhagen, Denmark

⁴Research Unit of General Practice, Department of Public Health, University of Southern Denmark, Odense, Denmark

⁵Department of Neurology, Slagelse Hospital, Slagelse, Denmark

⁶Department of Regional Health Research, University of Southern Denmark, Odense, Denmark

⁷Department of Environmental Science, Aarhus University, Roskilde, Denmark

⁸Global Centre for Clean Air Research (GCARE), University of Surrey, Guildford, United Kingdom

⁹Danish Big Data Centre for Environment and Health (BERTHA), Aarhus University, Roskilde, Denmark

¹⁰Research Unit for ORL – Head and Neck Surgery and Audiology, Odense University Hospital, Odense, Denmark

¹¹Department of Clinical Research, University of Southern Denmark, Odense, Denmark

¹²OPEN, Odense Patient data Explorative Network, Odense University Hospital, Odense, Denmark

¹³BRIDGE, Brain Research - Inter-Disciplinary Guided Excellence, Department of Clinical Research, Faculty of Health Sciences, University of Southern Denmark, Odense, Denmark

¹⁴Department of Natural Science and Environment, Roskilde University, Roskilde, Denmark

Contributors: MS, MLC, JHS, FBW, LW, ERP, OR-N, JD, and AHP conceived and designed the study and contributed to the methodology definition. JD, AHP, MS, and OR-N contributed to exposure assessment. MK and JK geocoded addresses and estimated air pollution. VHV geocoded addresses and estimated proportion of green space. MS, OR-N, and AHP acquired health and confounder data. MLC classified register data and conducted the statistical analysis. MLC and MS wrote the first draft of the manuscript. All authors commented and contributed to the interpretation of results and the final manuscript. The corresponding author attests that all listed authors meet the authorship criteria.

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Data sharing: As this study is based on data from the Danish national registers, which belong to the Danish Ministry of Health and Statistics Denmark, the authors are not allowed to share data in raw form.

Ethical approval: This study was approved by the local authorities (record No: 2018-DCRC-0055) and was conducted in accordance with principles of the Declaration of Helsinki. By Danish Law, the study does not require patients' consent and approval from ethical committees, as it is entirely based on data from the Danish national registers.

The lead author affirms (MLC) that the manuscript is an honest, accurate, and transparent account of the study being reported; that no important aspects of the study have been omitted; and that any discrepancies from the study as planned (and, if relevant, registered) have been explained.

Dissemination to participants and related patient and public communities: There are plans to disseminate the results of the research to the general public, mainly through media outreach (eg, press releases by the research institutions of the contributing authors, and plain language publications in popular and social media). Results will also be communicated to relevant clinicians in Denmark.

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Supplementary information: additional tables 1-8 and figures 1-6